

1        **Adaptive management in the context of barriers in European freshwater ecosystems**

2

3

4                                    *In Press in Journal of Environmental Management*

5

6

7                                    Kim Birnie-Gauvin<sup>1</sup>, Jeroen S. Tummers<sup>2</sup>, Martyn C. Lucas<sup>2</sup>, Kim Aarestrup<sup>1</sup>

8

9

10

11        <sup>1</sup> DTU AQUA, National Institute of Aquatic Resources, Section for Freshwater Fisheries

12        Ecology, Technical University of Denmark, Vejlsøvej 39, 8600 Silkeborg, Denmark

13        <sup>2</sup> Aquatic Animal Ecology Research Group, Department of Biosciences, University of Durham

14        South Road, Durham, DH1 3LE, UK

15

16

17

18

19        Author for correspondence: K. Birnie-Gauvin

20                                    [kbir@aquu.dtu.dk](mailto:kbir@aquu.dtu.dk)

21

22

23

24 **Abstract**

25 Many natural habitats have been modified to accommodate for the presence of humans and their  
26 needs. Infrastructures – such as hydroelectric dams, weirs, culverts and bridges – are now a  
27 common occurrence in streams and rivers across the world. As a result, freshwater ecosystems  
28 have been altered extensively, affecting both biological and geomorphological components of the  
29 habitats. Many fish species rely on these freshwater ecosystems to complete their lifecycles, and  
30 the presence of barriers, has been shown to reduce their ability to migrate and sustain healthy  
31 populations. In the long run, barriers may have severe repercussions on population densities and  
32 dynamics of aquatic animal species. There is currently an urgent need to address these issues  
33 with adequate conservation approaches. Adaptive management provides a relevant approach to  
34 managing barriers in freshwater ecosystems as it addresses the uncertainties of dealing with  
35 natural systems, and accommodates for future unexpected events, though this approach may not  
36 be suitable in all instances. A literature search on this subject yielded virtually no output. Hence,  
37 we propose a step-by-step guide for implementing adaptive management, which could be used to  
38 manage freshwater barriers.

39

40 **Keywords:** adaptive management, barriers, freshwater ecosystems, stakeholders, conservation

41

42

43

44

45

46

## 47 **1. Context: barriers in European freshwater ecosystems**

48 In comparison to their terrestrial counterparts, freshwater taxa are on average more imperiled  
49 (Dudgeon et al. 2006; Strayer and Dudgeon 2010; Carrizo et al. 2013). Freshwater fish species  
50 represent approximately 25% of all living vertebrates, many of which are threatened (IUCN  
51 2016). Given the linear nature of freshwater systems, connectivity may be heavily affected as a  
52 result of the presence of in-river barriers (Stanford et al. 1996). Historically, rivers and their  
53 surroundings have been used for anthropogenic purposes more than any other habitat, which over  
54 centuries, has led to the loss of the original integrity of water courses (Jungwirth 1998; Jager et  
55 al. 2001). Today, the majority of large rivers have been modified in one way or another – for the  
56 purposes of hydroelectric power plants (Welcomme 1995) or other artificial barriers like dams,  
57 weirs, or road crossings (Jungwirth et al. 2000; Nilsson et al. 2005), posing increasing threats to  
58 freshwater ecosystems and the mobile biota, particularly fish, that live within them (Arthington  
59 et al. 2016).

60 In Europe, all major rivers, except for the Pechora River in Russia (Studenov et al. 2008),  
61 are now fragmented by artificial dams and weirs (Tockner et al. 2009). The high (and increasing)  
62 density of river barriers is contributing to the poor habitat quality and loss of biodiversity of  
63 freshwater systems in contravention of the European Union's Water Framework Directive  
64 (Acreman and Ferguson 2010; Reyjol et al. 2014). Increasingly, barrier removal is viewed as a  
65 necessary management measure to reinstate natural connectivity within and amongst ecosystems  
66 (Garcia de Leaniz 2008; Tonra et al. 2015), though we still have little knowledge to make  
67 predictions about the biological and geomorphological trajectory of a river system once a barrier  
68 has been removed (Pizzuto 2002). Whilst removal projects for large barriers have revealed quick  
69 recovery of key biological components (Tonra et al. 2015), the same cannot be said of barriers in

70 small streams as evidence is currently lacking (Tummers et al. 2016a). The presence of small-to-  
71 medium sized impoundments (i.e., height below 10m) is extensive in European streams and  
72 rivers, providing us with every reason to investigate their effects in order to enhance and focus  
73 management efforts.

74

## 75 **2. Management of barriers**

76 Many barriers in European rivers originated in the 10<sup>th</sup> to 19<sup>th</sup> centuries to operate mills  
77 (Downward and Skinner 2005; Nützmänn et al. 2011) and a high proportion, often rebuilt or  
78 modified multiple times, are now redundant (Downward and Skinner 2005). However, some mill  
79 weirs are of historical significance or are being converted for operation as low-head  
80 hydroelectric power facilities (Watkin et al. 2012). Since the 1950s, the approach to implement  
81 dams for achieving water storage has been to design and operate reservoirs so that they fill with  
82 sediments slowly (Palmieri et al. 2001) but some are approaching the end of their operational  
83 lives. Currently, there are challenging issues regarding the proper management of barriers, which  
84 may be addressed by an adaptive management (AM) approach.

85         AM stems from the idea that ecosystem management and conservation practice is a  
86 dynamic process, and thus should be modified as we gain further knowledge to achieve  
87 management objectives (Holling 1978; Lindenmayer and Burgman 2005; Westgate et al. 2013).  
88 Such an approach is especially appropriate when dealing with ecological resources, which are  
89 dynamic in nature, and hence would provide an appropriate method to manage barriers (for  
90 example management of flow characteristics - see Baumgartner et al. 2014; Summers et al.  
91 2015). This dynamic conservation approach has grown greatly since the seminal work of Walters  
92 and Hilborn (1976) and Holling (1978), and is now considered fundamental to sustainable

93 practices (Westgate et al. 2013; Williams and Brown 2014). An adaptive approach requires  
94 extensive planning, along with an active and systematic effort to gather and document  
95 information, as well as the early involvement of stakeholders in the decision-making process  
96 (Lindenmayer and Burgman 2005). There are four fundamental elements to AM, as identified by  
97 Davis et al. 2001: (1) acknowledging the uncertainties associated with management policies, (2)  
98 formulating management policies as testable hypotheses, (3) searching, using and assessing  
99 information in order to test hypotheses, and (4) adapting management policies periodically as  
100 new information is acquired.

101         While AM is widely supported in theory (Fabricius and Cundill 2014), few real-world  
102 examples have been reported in practice (Keith et al. 2011; Westgate et al. 2013). Most  
103 applications test a single management option at a time, and change their approach only when it  
104 fails (Duncan and Wintle 2008; Keith et al. 2011). Our initial objective was to use a systematic  
105 approach to review the current state of research in adaptive barrier management of freshwater  
106 ecosystems. However, an all-time initial search on Web of Science using  
107 “(adaptiv\*)AND(manage\*)AND(freshwater)AND(barrier\*)” as the word string yielded only 17  
108 results, 13 of which were eliminated at the title level, and the remaining 4 were eliminated at the  
109 abstract level, suggesting that this area of research is highly understudied. We therefore opted to  
110 include a broader spectrum of literature, and gather relevant information on AM, in an attempt to  
111 apply it directly to barrier management in freshwater ecosystems. While we hoped to provide  
112 specific examples to demonstrate how AM has been successfully used in barrier management,  
113 the literature on the topic is scarce, although this is partly because some relevant projects that  
114 have adopted an AM ethos have not used this term explicitly (*Box 1*). Instead, we propose a step-  
115 by-step guide for how AM could be implemented in the management of freshwater barriers

116 (*Figure 1*), along with the potential benefits and challenges that come with using such an  
117 approach.

118

### 119 *Potential benefits*

120 One of the main advantages of AM is its regular reviews of the effectiveness and progress of the  
121 strategies currently in place in the river system being managed. Management objectives should  
122 be dynamic in natural systems, such as streams and rivers. Thus, as results are obtained (i.e.,  
123 research findings), objectives change, and accordingly, so should management strategies  
124 (exemplified in *Box 1*). Modelling tools are essential to understand how environmental factors  
125 may impact a system, and to predict the outcomes of various management options (Thom 2000;  
126 Bearlin et al. 2002). This approach helps to accommodate for future unexpected events by  
127 guiding the development of predictions and hypotheses, which is especially relevant in today's  
128 changing world. In barrier management, fish density, diversity, recruitment and spawning  
129 provide important metrics to track the efficacy of the management strategies currently in place.  
130 Regular revisions of these data will provide valuable information for modelling purposes and  
131 help promote future management success of barriers. Modelling is also beneficial to optimize an  
132 approach. In many ways, AM resembles a scientific experiment, where hypotheses are tested,  
133 and experimentation is carried out, thus rendering the conclusions to be drawn more robust  
134 (Linkov et al. 2004).

135

### 136 *Potential challenges*

137 A crucial component of AM is its ability to highlight the presence and importance of  
138 uncertainties, and to use these uncertainties when formulating and testing hypotheses to render

139 the process more efficient (Davis et al. 2001). In the context of AM, uncertainties arise from  
140 changing natural conditions, but also due to economic, social and political variability (Salwasser  
141 1993). Uncertainties must be managed by considering a wide range of adequate, realistic and  
142 reversible strategies - essentially replacing the uncertainty of a resource with the certainty of a  
143 process (Rodgers 1997). Results should be monitored continuously, and strategies adjusted as  
144 further knowledge is gained (Beese et al. 2003; Bunnell et al. 2003). While modeling is used to  
145 make predictions that take into account uncertainties, modeling with knowledge gaps (i.e., when  
146 all necessary information is not available) may exacerbate this uncertainty. AM is about  
147 “learning by doing”, and incorporating learnt lessons into future decisions (McDaniels and  
148 Gregory 2004). In the context of barriers, managers may use currently available findings (e.g., in  
149 the literature or reports) on the potential benefits of barrier removal (or the negative impacts of  
150 barrier implementation) for fish and apply this information to a new system, accepting alongside  
151 it the uncertainties that come with natural systems and populations.

152 In the real world, AM is difficult to attain successfully. Stakeholders may have  
153 conflicting perspectives despite a conservation objective agreed by all (Lindenmayer and  
154 Burgman 2005). In many instances, political and social circumstances make AM a difficult task  
155 to fulfill (*Table 1*). Scientists may not always recognize problems in AM sufficiently, as their  
156 solutions are not necessarily socially and politically acceptable (Salwasser 1993). A common  
157 caveat to AM is how it manages human motivation, often causing a source of problems in  
158 resources management (Ludwig et al. 1993), especially when the main concern should revolve  
159 around the resource itself. Stakeholders can sometimes be unwilling to compromise and/or  
160 accept *any* change, resulting in serious delays in management efforts, and may even completely  
161 stall the process. For example, dams are often constructed to alter flow regimes and generate

162 hydroelectricity (Dynesius and Nilsson 1994), causing substantial impacts on the ecological  
163 health of rivers (Bunn and Arthington 2002). Alternatively, old mills and weirs may have  
164 historical or cultural value to some, be used for recreational purposes (e.g., boating and fishing)  
165 and for supply of drinking water. Stakeholders from both sides must discuss management  
166 options, which will likely require compromises. In some cases minor stakeholders who remain  
167 completely unwilling to compromise or accept any form of change may simply have to be  
168 ignored.

169           When a resource collapses, all stakeholders typically agree that action must be taken.  
170 Nonetheless, complete consensus is almost unattainable, which puts management groups at a  
171 standstill. Some challenges are irreconcilable. We must therefore often take action before  
172 (scientific) consensus is reached. Unrealistic expectations can sometimes cause us to forget about  
173 the problem itself, but this adaptive approach is a trade-off between available data, and the need  
174 for immediate resource conservation. For example, the reinstatement of more natural conditions  
175 of streams and rivers via barrier removal may be a necessary action to conserve wild fish  
176 populations, despite the paucity of data on barrier removal.

177           Another challenge is that sometimes the problem is thought to be only marginal and so to  
178 initiate an AM process would be too costly and lengthy for the benefits. In this case, a potential  
179 solution may be to approach the entire river system as one management issue, rather than  
180 individual barriers within the system. In catchment management, barriers in small lowland  
181 streams are often disregarded and viewed as non-impactful obstacles, though their combined  
182 effects are in fact largely underestimated (Tummers et al. 2016a; Birnie-Gauvin et al. *in press*).  
183 In many instances, too much emphasis is placed on the measurable economic interests of  
184 stakeholders resulting in the underappreciation of conservation problems (often unmeasurable) at



185 hand, thereby slowing the process of experimentation, learning and adaptation. Management then  
186 becomes stuck at the modelling step because research is deemed too expensive, which comes at  
187 the cost of ecological sustainability.

#### 188 **4. Implementing adaptive management**

189 We propose a guide to implement adaptive management in the real world in *Figure 1*. Before  
190 initiating an AM approach, managers must first determine whether all of the four following  
191 components are present: (1) knowledge gaps, (2) prospects for learning and an expected  
192 ecological value, (3) opportunities for reconsiderations and alternative options (i.e., if only one  
193 option is viable, adaptive management is not an appropriate approach), and (4) sufficient  
194 funding. If all four components are present, then one may initiate the AM process, which begins  
195 with identifying and involving all relevant stakeholders. Managers must ask themselves three  
196 important questions: Are there highly valuable resources at stake? Is the scenario highly  
197 politically-involved? Is there a high degree of uncertainty revolving around this issue? If “yes” is  
198 answered to any of these questions, it is highly recommended that managers seek the help of  
199 independent peer-reviewers to help the decision-making process. The following step is one of the  
200 most critical steps in AM: setting clear objectives, which are agreed upon by all stakeholders.  
201 Without agreement, the process cannot move forward, sometimes at the cost of ecological  
202 resilience. Independent peer-reviewers may be helpful, but if the opinions of stakeholders are  
203 irreconcilable, then an alternate management approach must be investigated. Managers must then  
204 identify measurable indicators (of the chosen management actions), which must again be agreed  
205 upon. The modeling process subsequently begins, which helps the development of hypotheses  
206 and predictions, and vice versa. Following modeling, large-scale experimentation is carried out,  
207 where the outcomes are evaluated. If the outcomes are not satisfactory, then more modeling and

208 hypothesis-testing may be needed. If the outcomes are deemed satisfactory by stakeholders, the  
209 agreed upon management actions may be implemented and evaluated repeatedly at regular  
210 intervals. Discussions, reflections and adaptations to the management approach should be  
211 undertaken continuously. Every step of this process should be documented adequately.

212

## 213 **5. Conclusion and an outlook to the future**

214 In many cases, “we know too little about how threats operate at large scales to be able to prevent  
215 or mitigate them” (Abell 2002). Adaptive management attempts to deal with the uncertainties  
216 that come with “knowing too little”. Nonetheless, there are instances in which adaptive  
217 management is simply not an acceptable option (*Table 2*), a fact which cannot be understated -  
218 adaptive management is by no means the answer to every conservation issue. There exist several  
219 guidelines and prerequisites that must be met before one can set out to implement an adaptive  
220 management approach (*Figure 1*). Under certain circumstances, it may be valuable to combine  
221 an adaptive management approach with other approaches to developed tools which can be  
222 applied at a wider scale (e.g., Fuzzy Cognitive Mapping, Özesmi et al. 2004). In cases when  
223 adaptive management can be used, it is important that the process and outcomes - for both  
224 failures and successes - be documented (either as a report or peer-reviewed article) so that others  
225 can benefit from it. It may also be beneficial to managers if a formal framework on how to  
226 implement adaptive management is available.

227

## 228 **Acknowledgements**

229 This contribution was funded by the European Union AMBER (Adaptive Management of  
230 Barriers in European Rivers) project as part of the Horizon 2020 Framework Programme.

231

232

233 **References**

234 Abell, R. (2002). Conservation biology for the biodiversity crisis: a freshwater follow-up.  
235 *Conservation Biology* **16**: 1435-1437.

236

237 Acreman, M.C., Ferguson, A.J.D. (2010). Environmental flows and the European Water  
238 Framework Directive. *Freshwater Biology* **55**: 32-48.

239

240 Arthington, A.H., Dulvy, N.K., Gladstone, W., Winfield, I.J. (2016). Fish conservation in  
241 freshwater and marine realms: status, threats and management. *Aquatic Conservation: Marine  
242 and Freshwater Ecosystems* **26**: 838–857.

243

244 Baumgartner, L.J., Conallin, J., Wooden, I., Campbell, B., Gee, R., Robinson, W.A., Mallen-  
245 Cooper, M. (2014). Using flow guilds of freshwater fish in an adaptive management framework  
246 to simplify environmental flow delivery for semi-arid riverine systems. *Fish and Fisheries* **15**:  
247 410-427.

248

249 Bearlin, A.R., Schreiber, E.S.G., Nicol, S.J., Starfield, A.M., Todd, C.R. (2002). Identifying the  
250 weakest link: simulating adaptive management of the reintroduction of a threatened fish.  
251 *Canadian Journal of Fisheries and Aquatic Sciences* **59**: 1709-1716.

252

253 Beese, W. J., Dunsworth, B. G., Zielke, K., Bancroft, B. (2003). Maintaining attributes of old-  
254 growth forests in coastal BC through variable retention. *The Forestry Chronicle* **79**: 570-578.

255

256 Birnie-Gauvin, K., Aarestrup, K., Riis, T.M.O., Jepsen, N., Koed, A. (in press). Shining the light  
257 on the loss of rheophilic fish habitat in lowland rivers as a forgotten consequence of barriers and  
258 its implications for management. *Aquatic Conservation: Marine and Freshwater Ecosystems*.  
259 doi: 10.1002/aqc.2795

260

261 Bracken, F.S.A., Lucas, M.C. (2013). Potential impacts of small-scale hydroelectric power  
262 generation on downstream moving lampreys. *River Research and Applications* **29**: 1073-1081.

263

264 Bracken, F.S.A., Hoelzel, A.R., Hume, J.B., Lucas, M.C. (2015). Contrasting population genetic  
265 structure among freshwater-resident and anadromous lampreys: the role of demographic history,  
266 differential dispersal, and anthropogenic barriers to movement. *Molecular Ecology* **24**: 1188-  
267 1204.

268

269 Bunn, S.E., Arthington, A.H. (2002). Basic principles and ecological consequences of altered  
270 flow regimes for aquatic biodiversity. *Environmental Management* **20**: 492-507.

271

272 Bunnell, F.L., Dunsworth, B.G. (2004). Making adaptive management for biodiversity work the  
273 example of Weyerhaeuser in coastal British Columbia. *The Forestry Chronicle* **80**: 37-43.

274

275 Carrizo, S.F., Smith, K.G., Darwall, W.R.T. (2013). Progress towards a global assessment of the  
276 status of freshwater fishes (Pisces) for the IUCN Red List: application to conservation  
277 programmes in zoos and aquariums. *International Zoo Yearbook* **47**: 46-64.  
278

279 Davis, M.B., Shaw, R.G. (2001). Range shifts and adaptive responses to quaternary climate  
280 change. *Science* **292**: 673-679.  
281

282 Downward, S., Skinner, K. (2005). Working rivers: the geomorphological legacy of English  
283 freshwater mills. *Area* **37**: 138-147.  
284

285 Dudgeon, D., Arthington, A.H., Gessner, M.O., Kawabata, Z.I., Knowler, D.J., Lévêque, C.,  
286 Sullivan, C.A. (2006). Freshwater biodiversity: importance, threats, status and conservation  
287 challenges. *Biological Reviews* **81**: 163-182.  
288

289 Duncan, D.H., Wintle, B.A. (2008). Towards adaptive management of native vegetation in  
290 regional landscapes. In: Pettit, C., Bishop, I., Cartwright, W., Duncan, D., Lowell, K., Pullar, D.  
291 eds. *Landscape Analysis and Visualisation. Spatial models for Natural Resource Management*  
292 *and Planning*. Berlin, Springer-Verlag GmbH, pp. 159-182.  
293

294 Dynesius, M., Nilsson, C. (1994). Fragmentation and flow regulation of river systems in the  
295 northern third of the world. *Science* **5186**: 753-762.  
296

297 Fabricius, C., Cundill, G. (2014). Learning in adaptive management: insights from published  
298 practice. *Ecology and Society* **19**: 29-36.  
299

300 Foulds, W.L., Lucas, M.C. (2013). Extreme inefficiency of two conventional technical fishways  
301 used by European river lamprey (*Lampetra fluviatilis*). *Ecological Engineering* **58**: 423-433.  
302

303 Garcia de Leaniz, C. (2008). Weir removal in salmonid streams: implications, challenges and  
304 practicalities. *Hydrobiologia* **609**: 83-96.  
305

306 Holling, C.S. (1978). *Adaptive Environmental Management and Assessment*. Chichester, John  
307 Wiley and Sons.  
308

309 IUCN (2016). IUCN Red List: An overview of the IUCN red list. International Union for the  
310 Conservation of Nature and Natural Resources.  
311

312 Jager, H.I., Chandler, J.A., Lepla, K.B., Winkle, W.V. (2001). A theoretical study of river  
313 fragmentation by dams and its effects on white sturgeon populations. *Environmental Biology of*  
314 *Fishes* **60**: 347-361.  
315

316 Jang, M.--H., Lucas, M.C. (2005). Reproductive ecology of the river lamprey. *Journal of Fish*  
317 *Biology* **66**: 499-512.  
318

319 Jungwirth, M. (1998). River continuum and fish migration – Going beyond the longitudinal river  
320 corridor in understanding ecological integrity. In: Jungwirth, M., Schmutz, M. S., Weiss, S.

321 (eds.). Fish migration and fish bypasses. Fishing News Books. Blackwell Science Ltd., Oxford,  
322 pp. 127-145.  
323  
324 Jungwirth, M., Muhar, S., Schmutz, S. (2000). Fundamentals of fish ecological integrity and  
325 their relation to the extended serial discontinuity concept. *Hydrobiologia* **422**: 85-97.  
326  
327 Keith, D.A., Martin, T.G., McDonald-Madden, E., Walters, C. (2011). Uncertainty and adaptive  
328 management for biodiversity conservation. *Biological Conservation* **144**: 1175-1178.  
329  
330 Lindenmayer, D.H., Burgman, M.A. (2005). Practical Conservation Biology. Collingwood,  
331 CSIRO. 610 pp.  
332  
333 Linkov, T.S., Varghese, A., Jamil, S., Seager, T.P., Kiker, G., Bridges, T. (2004). Multi-criteria  
334 decision analysis: framework for applications in remedial planning for contaminated sites. In:  
335 Linkov, I., Ramadan, A. (eds.). Comparative Risk Assessment and Environmental Decision  
336 Making. Amsterdam, Kluwer, pp. 15-54.  
337  
338 Lucas, M.C., Bubb, D.J., Jang, M.-H., Ha, K., Masters, J.E.G. (2009). Availability of and access  
339 to critical habitats in regulated rivers: iImpacts of low-head barriers on threatened lampreys.  
340 *Freshwater Biology* **54**: 621-634.  
341  
342 Ludwig, D. (1993). Environmental sustainability: magic, science, and religion in natural resource  
343 management. *Ecological Applications* **3**: 555-558.  
344  
345 McDaniels, T.L., Gregory, R. (2004). Learning as an objective within a structural risk  
346 management decision process. *Environmental Science and Technology* **38**: 1921-1926.  
347  
348 Nilsson, C., Reidy, C.A., Dynesius, M., Revenga, C. (2005). Fragmentation and flow regulation  
349 of the world's large river systems. *Science* **308**: 405-408.  
350  
351 Nunn, A.D., Harvey, J.P., Noble, R.A.A., Cowx, I.G., (2008). Condition assessment of lamprey  
352 populations in the Yorkshire Ouse catchment, north-east England, and the potential influence of  
353 physical migration barriers. *Aquatic Conservation: Marine and Freshwater Ecosystems* **18**: 175-  
354 189.  
355  
356 Nützman, G., Wolter, C., Venohr, M., Pusch, P. (2011). Historical patterns of anthropogenic  
357 impacts on freshwaters in the Berlin-Brandenburg region. *Die Erde* **142**: 41-64.  
358  
359 Özesmi, U., Özesmi, S. L. (2004). Ecological models based on people's knowledge: a multi-step  
360 fuzzy cognitive mapping approach. *Ecological Modelling* **176**: 43-64.  
361  
362 Ovidio, M., Philippart, J.C. (2002). The impact of small physical obstacles on upstream  
363 movements of six species of fish. *Hydrobiologia* **483**: 55-69.  
364  
365 Palmieri, A., Shah, F., Dinar, A. (2001). Economics of reservoir sedimentation and sustainable  
366 management of dams. *Journal of Environmental Management* **61**: 149-163.

367  
368 Pizzuto, J. (2002). Effects of dam removal on river form and process: Although many well-  
369 established concepts of fluvial geomorphology are relevant for evaluating the effects of dam  
370 removal, geomorphologists remain unable to forecast stream channel changes caused by the  
371 removal of specific dams. *BioScience* **52**: 683-691.  
372  
373 Reyjol, Y., Argillier, C., Bonne, W., Borja, A., Buijse, A.D., Cardoso, A.C., Prat, N. (2014).  
374 Assessing the ecological status in the context of the European Water Framework Directive:  
375 Where do we go now? *Science of the Total Environment* **497**: 332-344.  
376  
377 Rodgers, W.H. (1997). Environmental Law, second edition. West Publishing, St. Paul,  
378 Minnesota, USA.  
379  
380 Royal Haskoning (2010). Yorkshire Derwent Restoration Plan.  
381 [http://www.therrc.co.uk/DesignatedRivers/Yorks\\_Derwent\\_Restoration\\_Plan.pdf](http://www.therrc.co.uk/DesignatedRivers/Yorks_Derwent_Restoration_Plan.pdf)  
382  
383 Salwasser, H. (1993). Sustainability needs more than better science. *Ecological Applications* **3**:  
384 587-589.  
385  
386 Saunders, D.A., Hobbs, R.J., Margules, C.R. (1991). Biological consequences of ecosystem  
387 fragmentation: A review. *Conservation Biology* **5**: 18-32.  
388  
389 Silva, S., Lowry, M., Macaya, C., Byatt, B., Lucas, M.C. (2017). Can navigation locks be used to  
390 help migratory fishes with poor swimming performance pass tidal barrages? A test with  
391 lampreys. *Ecological Engineering* **102**: 291-302.  
392  
393 Stanford, J. A., Ward, J. V., Liss, W. J., Frissell, C. A., Williams, R. N., Lichatowich, J. A.,  
394 Coutant, C. C. (1996). A general protocol for restoration of regulated rivers. *River Research and*  
395 *Applications* **12**: 391-413.  
396  
397 Strayer, D.L., Dudgeon, D. (2010). Freshwater biodiversity conservation: recent progress and  
398 future challenges. *Journal of the North American Benthological Society* **29**: 344-358.  
399  
400 Studenov, I., Antonova, V., Chuksina, N., Titov, S. (2008). Atlantic salmon of the Pechora  
401 River. SevPINRO, Arkhangelsk, pp 52.  
402  
403 Summers, M.F., Holman, I.P., Grabowski, R.C. (2015). Adaptive management of river flows in  
404 Europe: A transferable framework for implementation. *Journal of Hydrology* **531**: 696-705.  
405  
406 Thom, R.H. (2000). Adaptive management of coastal ecosystem restoration projects. *Ecological*  
407 *Engineering* **15**: 365-372.  
408  
409 Tockner, K., Uehlinger, U., Robinson, C.T. (eds.) (2009). Rivers of Europe. Academic Press,  
410 Oxford, UK.  
411

412 Tonra, C.M., Sager-Fradkin, K., Morley, S.A., Duda, J.J., Marra, P.P. (2015). The rapid return of  
413 marine-derived nutrients to a freshwater food web following dam removal. *Biological*  
414 *Conservation* **192**: 130-134.

415  
416 Travade, F., Larinier, M. (1992). Downstream migration: problems and facilities. *Bulletin*  
417 *Francais de la Peche et de la Pisciculture* **364**: 181-207.

418  
419 Tummers, J.S., Hudson, S., Lucas, M.C. (2016a). Evaluating the effectiveness of restoring  
420 longitudinal connectivity for stream fish communities: towards a more holistic approach. *Science*  
421 *of the Total Environment* **569-570**: 850-860.

422  
423 Tummers, J.S., Winter, E., Silva, S., O'Brien, P., Jang, M.-H., Lucas, M.C. (2016b). Evaluating  
424 the effectiveness of a Larinier super active baffle fish pass for European river lamprey *Lampetra*  
425 *fluviatilis* before and after modification with wall-mounted studded tiles. *Ecological Engineering*  
426 **91**: 183-194.

427  
428 Vowles, A.S., Don, A.M., Karageorgopoulos, P., Kemp, P.S. (2017). Passage of European eel  
429 and river lamprey at a model weir provisioned with studded tiles. *Journal of Ecohydraulics* **2**:  
430 DOI 10.1080/24705357.2017.1310001

431  
432 Walters, C.J., Holling, C.S. (1990). Large-scale management experiments and learning by doing.  
433 *Ecology* **71**: 2060-2068.

434  
435 Walters, C., Hillborn, R. (1976). Adaptive control of fishing systems. *Journal of the Fisheries*  
436 *Research Board of Canada* **33**: 145-159.

437  
438 Watkin, L.J., Kemp, P.S., Williams, I.D., Harwood, I.A. (2012). Managing sustainable  
439 development conflicts: the impact of stakeholders in small-scale hydropower schemes.  
440 *Environmental Management* **49**: 1208-1223.

441  
442 Welcomme, R.L. (1995). Relationships between fisheries and integrity of river systems.  
443 *Regulated Rivers: Research and Management* **11**: 121-136.

444  
445 Westgate, M.J., Likens, G.E., Lindenmayer, D.B. (2013). Adaptive management of biological  
446 systems: a review. *Biological Conservation* **158**: 128-139.

447  
448 Williams, B.K., Brown, E.D. (2014). Adaptive management: from more talk to real  
449 action. *Environmental Management* **53**: 465-479.

450  
451  
452  
453  
454

455 *Table 1.* Stakeholders and their incentive for barrier management.  
 456

<b>Stakeholder</b>	<b>Underlying incentive</b>
Hydroelectric dam owner	Economic value, provision of energy
Residents of local municipality	Flood risk (economic impact), cultural heritage, recreation (boating, fishing, wildlife)
Environmental protection agencies	Flow gauging, flood risk
Water companies	Economic value, water abstraction for drinking water
Farmers of adjacent land	Economic value, water abstraction for crops, flood risk adjacent to river
Boat navigation	Channel depth management, economic value
Highways / rail authority	Economic value, transport where barrier issue is linked to road/rail transport (culvert, bridge infrastructure)
Fish farmers	Economic value, stocking
Recreational fishing	Economic value, intrinsic values
Commercial fishing	Economic value, food provision
Conservation bodies	Maintaining biodiversity, environmental and population sustainability

457  
 458  
 459  
 460  
 461  
 462  
 463  
 464  
 465  
 466  
 467  
 468  
 469  
 470  
 471  
 472  
 473  
 474  
 475



476 *Table 2.* Limitations of the adaptive management approach.  
477

**Instances when NOT to use adaptive management**

---

To delay a process.

---

When there are no knowledge gaps.

---

When no clear objectives have been set.

---

When funding is a problem.

---

When opportunities for improvement lack.

---

When later reconsiderations are not an option.

---

When alternatives are limited.

---

When mistakes are irreversible.

---

When no measurable indicators are available.

---

Irreconcilable stakeholders

---

478  
479  
480

481 **Box 1.** Adaptive management of river barriers in action - a case study

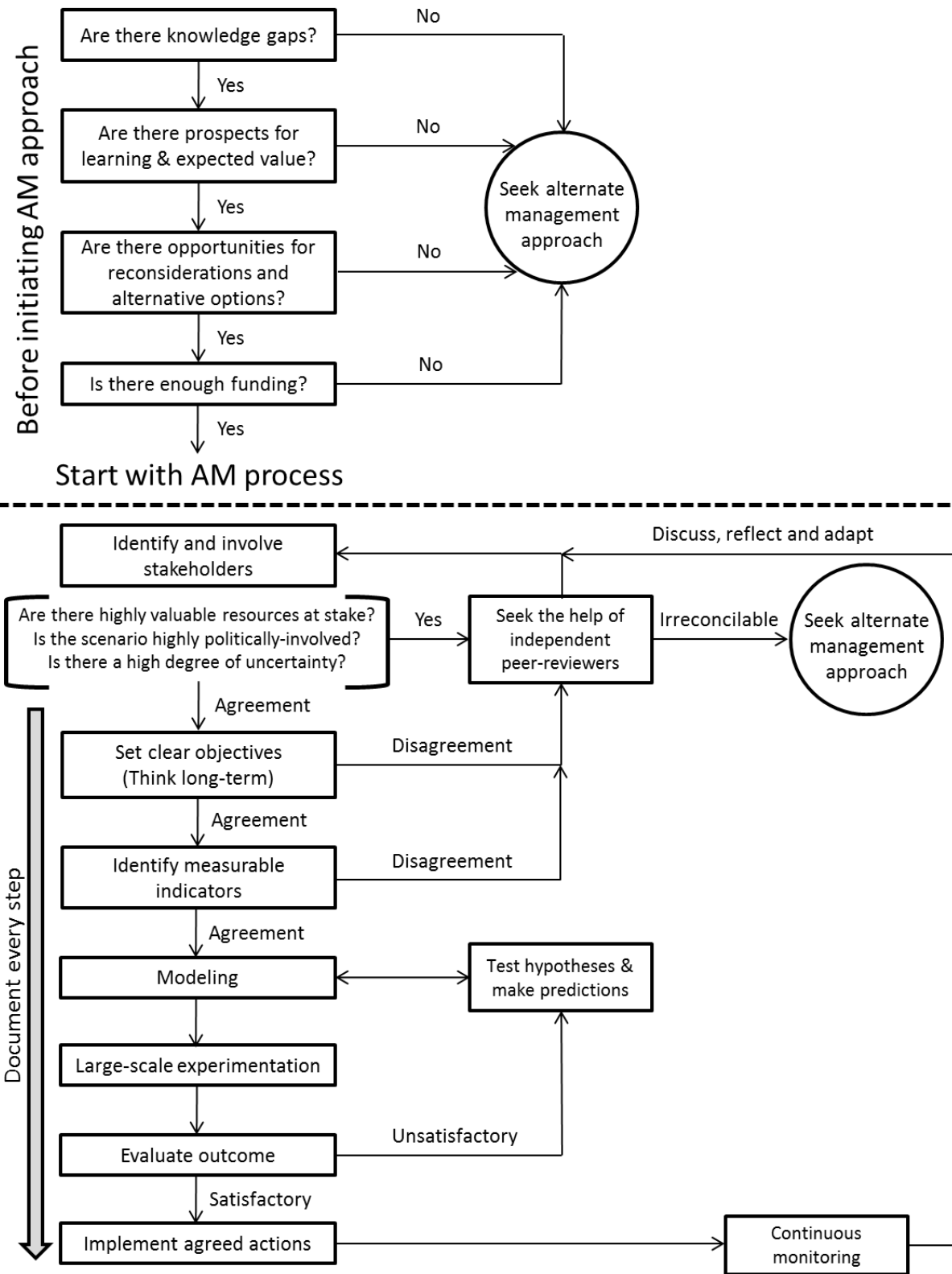
482

483 The Yorkshire Derwent, northeast England is a tributary of the Humber, the UK's largest drainage. The  
484 Derwent catchment is mostly rural and has good water quality, suitable for potable supply after treatment.  
485 The catchment runs off the North Yorkshire Moors but the last 75 km of river falls only 20 m (mostly at  
486 six river barriers), creating a large managed floodplain. The downstream-most 35km of this comprises  
487 herb-rich damp meadows. From km 68 to the confluence with the Humber, the river was designated a  
488 national Site of Special Scientific Interest (SSSI) in 1975 and an EU Special Area of Conservation (SAC)  
489 in 2005. Adjacent wetlands form an EU Special Protection Area (SPA) for wetland birds and a RAMSAR  
490 wetland site. *Ranunculion fluitantis* / *Callitrichio-Batrachion* habitat and river lamprey *Lampetra*  
491 *fluviatilis* were primary reasons for selection of the lower Derwent as an SAC. However, since 2003,  
492 Natural England (NE) determined the Derwent SAC to be in unfavourable condition for these features.  
493 Key pressures were identified as siltation, and in-river barriers to fish movement. Additional management  
494 issues relating to River Derwent barriers are flood risk management (towns along the lower Derwent have  
495 flooded multiple times in recent decades); potable water supply (the lower two barriers stabilise water  
496 levels upstream for abstraction to 5 million people); new low-head hydroelectricity (the Environment  
497 Agency [EA] is required to support renewable power development alongside its environmental protection  
498 duties); flow-gauging (EA gauges river flow from several weirs) and navigation (on the lower 35 km of  
499 river, including to and from the Humber, via Barmby tidal barrage, the downstream-most barrier,  
500 managed by EA). In 2003 the EA and NE sought to develop a long-term ecological restoration plan for  
501 the river (River Derwent Restoration Project, RDRP), in an adaptive framework and consulted with a  
502 wide range of stakeholders, identifying objectives and information needs.

503

504 To provide information for the RDRP and more widely, lamprey research on the Derwent has included  
505 determining their abundance and distribution (Jang and Lucas 2005; Nunn et al. 2008; Lucas et al. 2009);  
506 the distribution and use of lamprey habitats (Jang and Lucas 2005); the effect of habitat fragmentation on  
507 lamprey population genetics (Bracken et al. 2015); migration and passability of different barriers and the  
508 utility of various fishway designs (Lucas et al. 2009; Foulds and Lucas 2013; Tummers et al. 2016b; Silva  
509 et al. 2017); and hydroelectricity impacts on lampreys (Bracken and Lucas 2013). The River Derwent  
510 Restoration Plan (Royal Haskoning 2010) evaluated multiple options for solving in-river barrier impacts,  
511 site by site, including full barrier removal, barrier height reduction and provision of fishways. These  
512 options were appraised in concert with opportunities for reducing flood risk, managing key infrastructure  
513 (e.g. water abstraction), supporting hydroelectricity development, and the economic costs and benefits.  
514 This continues to be an ongoing adaptive process. For example, in 2010 EA decided not to remove its  
515 redundant flow-gauging weir at rkm 40, but to allow commercial hydroelectric development there and  
516 build a Larinier superactive baffle fishway, in the expectation that this would be usable by river lamprey.  
517 Research has since shown the Larinier design to be ineffective for lamprey upstream passage (Tummers  
518 et al. 2016b) and alternative passage solutions are being researched (Vowles et al. 2017). Modelling of  
519 weir height reductions at several other sites has been done and engineering options and costs for height  
520 reduction are actively being pursued. Since 2006, at Barmby tidal barrage, operations and automated  
521 controls have been altered, tested and improved to enhance fish passage, particularly through the use of  
522 the navigation lock in 'fishway mode' (Silva et al. 2017). Although this is intended for lamprey migration  
523 it can likely benefit eels, flatfish and Atlantic salmon *Salmo salar*, which are starting to recolonize the  
524 river after an absence of many decades due mostly to pollution of the Humber estuary.

525 *Figure 1.* Proposed step-by-step guide to implement an adaptive approach in barrier  
 526 management.  
 527



528